

## PROGRAMA DE PÓS-GRADUAÇÃO EM OCEANOGRAFIA AMBIENTAL UNIVERSIDADE FEDERAL DO ESPÍRITO SANTO

## UNIVERSIDADE FEDERAL DO ESPÍRITO SANTO CENTRO DE CIÊNCIAS HUMANAS E NATURAIS PROGRAMA DE PÓS-GRADUAÇÃO EM OCEANOGRAFIA AMBIENTAL

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Efeitos climáticos em assembleias macrobentônicas de um estuário tropical da Ecorregião marinha Leste do Brasil

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Dissertação apresentada ao Programa de Pós-Graduação em Oceanografia Ambiental da Universidade Federal do Espírito Santo, como requisito parcial para obtenção do título de Mestre em Oceanografia Ambiental. Orientador: Prof. Dr. Angelo Fraga Bernardino Luiz Eduardo de Oliveira Gomes

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## **RESUMO**

Espera-se que as alterações climáticas e a intensificação das inundações e secas alterem significativamente os ecossistemas marinhos. Compreender os efeitos de secas e inundações nos estuários é de extrema importância para compreender os impactos climáticos potenciais em assembleias bentônicas. O estuário Piraque-Açú-Mirim é um ecossistema relativamente bem preservado localizado na Ecorregião Marinha do Leste Brasileira, e foi utilizado como um estudo de caso para a compreensão de variações sazonais e sua influência sobre as assembleias bentônicas. O monitoramento do estuário do Piraquê-Açú-Mirim é um dos mais extensos nos estuários brasileiros (n = 14 meses), sendo vantajoso pois o estuário está em uma área com evidente decréscimo na precipitação média anual durante a última década. Aqui apresentamos um monitoramento temporal (meses) de precipitação, salinidade e temperatura da água, e sedimentos na região estuarina, cujo objetivo foram i) identificar mudanças temporais em sedimentos e assembleias bentônicas no estuário Piraque-Açú-Mirim; e ii) relacionar a influência potencial do evento de seca às propriedades sedimentares e da macrofauna. Nossa hipótese é que as assembleias bentônicas seriam diferentes ao longo do tempo e entre as zonas halinas respondendo a mudanças na salinidade e temperatura da água. As zonas halinas foram definidas de acordo com dados ambientais anteriores do estuário e coletas pilotos feitas com data logger de condutividade/temperatura (OnSet) e multiparâmetro. Um evento de El-Niño ocasionou um intenso período seco entre os anos de 2015-2016, com drástica redução nas taxas de precipitação anual durante um dos mais fortes El-Niño registrados desde 1950. A salinidade e a temperatura em cada zona halina (euhalina e polihalina) entre as marés de sizígia e quadratura foram diferentes. O padrão temporal de salinidade e temperatura das zonas euhalina e polihalina do estuário Piraquê-Açú-Mirin modificaram entre os períodos de Seca e Chuvoso no ano de 2015 se comparados com o período de Seca em 2016, resultando em menores oscilações diárias na salinidade e temperatura estuarina. O conteúdo de lama ao longo do ano de 2015 foi significativamente menor que os observados durante o período de Seca em 2016 (~ 100%). A clorofila a nas estações Seca e Chuvoso 2015 foi significativamente maior que durante a estação Seca 2016. A intrusão da água marinha nas zonas polihalinas no estuário Piraquê-Açú-Mirim provavelmente contribuiu para o aumento do teor de lama nas zonas euhalina e polihalina. Durante o monitoramento do estuário do Piraquê-Açú-Mirim as assembleias bentônicas nas zonas euhalina e polihalina foram dominadas pelos mesmos poucos táxons (Magelonidae, Sternaspidae e Capitellidae). Porém, foram observadas mudanças na composição dos demais táxons menos abundantes e provavelmente menos tolerantes. Dentre as variáveis sedimentares, o teor de lama foi o principal fator que contribuiu para as mudanças temporais das assembleias bentônicas, e provavelmente evidencia mudanças no perfil salino estuarino. Temperatura da água esteve relacionada a mudanças na riqueza e equitabilidade J' das assembleias bentônicas, e com mudanças na composição das assembleias do estuário Piraquê-Açú-Mirim. Nossos resultados corroboram que a temperatura influencia mudanças nas assembleias bentônicas durante a seca, bem como a diminuição da diversidade e densidade de taxons dominantes no estuário Piraquê-Açú-Mirim. Ainda são necessários estudos para melhor compreender a relação entre as variáveis dependentes climáticas (ex. salinidade e temperatura) e assembleias bentonicas, bem como os fatores que controlam sua variabilidade. Projetos de monitoramento de longo prazo ainda são necessários, especialmente em países subdesenvolvidos e em desenvolvimento. No caso do Brasil, estes estudos, são essenciais para melhor gerir importantes sistemas, como o estuário do Piraquê-Açú-Mirim e as baías de Todos os Santos e Paranaguá versus as tendências causadas pelo iminente aumento da intensidade e freqüência das secas no século atual.

Palavras-chave: Redução pluviométrica, Monitoramento, Ecologia bêntica, Mudanças climáticas

## ABSTRACT

Climate change and the intensification of floods and droughts events are expected to significantly alter marine ecosystems. Understand the effects of droughts and floods on estuaries are of utmost importance to understanding the potential climate impacts on estuaries. The Piraque-Açú-Mirim estuary is a relatively well-preserved ecosystem located in the Eastern Brazilian Marine Ecoregion, and was used as a case study for the understanding of seasonal variations and their influence on benthic assemblages. The monitoring of the Piraquê-Açú-Mirim estuary is one of the most extensive of Brazilian estuaries (n=14 months) and advantageous because the estuary is in an area with evidence of significant decrease in annual average rainfall over the last decade. Here we present a temporal (months) monitoring of precipitation, salinity and temperature of water, and sediments in the estuarine region, whose objective is i) identify temporal changes in sediments and benthic assemblages in the Piraque-Açú-Mirim estuary; and ii) relate the potential influence of the drought event to the sedimentary and macrofaunal properties. Our hypothesis is that macrofaunal assemblages would be different over time and between haline zones responding to changes in salinity and water temperature. Haline zones were defined according to previous environmental data from the estuary and pilot samples with the data logger conductivity/temperature (OnSet) and multiparameter. An El Niño event caused an intense dry season between the years 2015-2016, with drastic reductions in annual precipitation rates during one of the strongest El-Niño recorded since 1950. Salinity and temperature at each haline zone (euhaline and polyhaline) were different between spring and neap tide tides. Temporal pattern of salinity and temperature in the Piraquê-Açú-Mirin estuary changes from dry and wet seasons of 2015 compared to dry season of 2016 resulting in lower daily oscillations in salinity and temperature of the estuary. Mud content throughout the year 2015 was significantly lower than those observed during the Drought season in 2016 (~ 100%). Chlorophyll-a in dry and wet seasons of 2015 were significantly higher than during dry season of 2016. Marine water intrusion in polyhaline zones in the Piraquê-Açú-Mirim estuary likely contributed to the increase of mud content at both euhaline and polyhaline zones. During the monitoring of the Piraquê-Açú-Mirim estuary the benthic assemblages in euhaline and polyhaline zones were dominated by the same few taxa (Magelonidae, Sternaspidae and Capitellidae). However, changes in the composition of the other less abundant and probably less tolerant taxa were observed. Mud content was the main factor that contributed to the temporal changes of the benthic assemblages, and probably shows changes in the estuarine saline profile. Water temperature was related to changes in richness S and Pielou J', and to changes in the composition of the assemblages of the Piraquê-Acú-Mirim estuary. Our results corroborate that temperature influence changes in benthic assemblages during drought, as well as the reduction of diversity and density of dominant taxa in the Piraquê-Açú-Mirim estuary. Further studies are need to better understand the relationship between climatic dependent variables (e.g. salinity and temperature) and benthic assemblages, as well as the factors controlling their variability. Long-term monitoring projects are still required, especially in underdeveloped and developing countries. In the case of Brazil, these studies are essential to better manage important systems, such as the Piraquê-Açú-Mirim estuary and the bays of Todos os Santos and Paranaguá versus the trends caused by the imminent increase in intensity and frequency of droughts in the current century.

Keywords: Rainfall decrease, Monitoring, Benthic ecology, Climate change

## LISTA DE TABELAS

Tabela 1. Variação do pH, Total de sólidos dissolvedos (TDS) e Oxigênio dissolvedo (DO) nas zonas euhalina e polihalina entre as estações seca e chuvoso de 2015 e seca de 2016 no estuário Piraquê-Açú-Mirim.

Tabela 2. Densidade média (ind.m<sup>2</sup>) e abundância relativa das assembleias macrobentônicas nas zonas euhalina e polihalina entre as estações seca e chuvoso de 2015 e seca de 2016 no estuário Piraquê-Açú-Mirim. = cinco melhores ranqueadas. P = Polichaeta, O = Oligochaete, M = Mollusca, C = Crustacea.

Tabela 3. Modelo linear baseado em distância (DistLM) da similaridade de Bray-Curtis em assembleias macrobentônicas e propriedades sedimentares (conteúdo de lama, Conteúdo orgânico total, Carbonato de cálcio, Clorofila a e Pheopigmentos) entre as estações seca e chuvoso de 2015 e seca de 2016 no estuário Piraquê-Açú-Mirim. Negrito = p significativo.

Tabela 4. Melhores correlações das assembleias macrobentônicas e propriedades sedimentares (conteúdo de lama, Conteúdo orgânico total, Carbonato de cálcio, Clorofila a e Pheopigmentos) no estuário Piraquê-Açú-Mirim. Pw - coeficientes ponderados de Spearman.

Tabela 5. Sítios, país e duração do estudo de monitoramento com eventos em outros estuários em relação ao estuário Piraquê-Açú-Mirim. Total = Total de eventos de amostragem.

Tabela S1. Variação no conteúdo de lama (%), total de conteúdo orgânico (%), Carbonato de cálcio (%), Clorofila *a* ( $\mu$ g.g-1) e feopigmentos ( $\mu$ g.g-1) através dos meses de monitoramento (maio de 2015 a junho de 2016, 14 meses) nas zonas euhalina e polihalina do estuário Piraquê-Açú-Mirin.

Tabela S2. Variação na riqueza S, densidade (ind.m<sup>2</sup>), Pielou J', Shannon H' (loge) e biomassa média (g.m<sup>2</sup>) através dos meses de monitoramento (maio de 2015 a junho de 2016, 14 meses) nas zonas euhalina e polihalina do estuário Piraquê-Açú-Mirin.

Tabela S3. Média da densidade e abindância relativa da macrofauna bentônica através dos meses de monitoramento (maio de 2015 a junho de 2016, 14 meses) nas zonas euhalina e polihalina do estuário Piraquê-Açú-Mirin. = top ranked macrofaunal taxa. P = Poliqueta, O = Oligoqueta, M = Molusco, C = Crustaceo.

## LISTA DE FIGURAS

Figura 1. Mapa do estuário Piraquê-Açú-Mirim evidenciando os sítios amostrais.

Figura 2. Superior: Média da precipitação anual de 1948 a 2016 (estações climáticas ANA e UFES; 1948 a 2016) do estuário Piraquê-Açú-Mirim; Inferior: e precipitação mensal total durante o estudo (estações climáticas ANA e UFES; 2015 a 2016) e informações pretéritas (estações climáticas ANA; 1948 a 2014) no estuário Piraquê-Açú-Mirim.

Figura 3. Salinidade durante a maré de sizígia e quadratura durante os dias de monitoramento nas estações seca e chuvosa de 2015 e seca de 2016vnas zonas euhalina e polihalina do estuário at Piraquê-Açú-Mirim. Dias de monitoramento nas zonas euhalina (40 dias, estação seca de 2015 (6 dias), estação chuvosa de 2015 (24 dias) e estação seca de 2016 (10 dias)); e zona polihalina (35 dias; estação seca de 2015 (5 dias), estação chuvosa de 2015 (22 dias) e estação seca de 2016 (8 dias)).

Figura 4. Temperatura (°C) durante a maré de sizígia e quadratura durante os dias de monitoramento nas estações seca e chuvosa de 2015 e seca de 2016vnas zonas euhalina e polihalina do estuário at Piraquê-Açú-Mirim. Dias de monitoramento nas zonas euhalina (40 dias, estação seca de 2015 (6 dias), estação chuvosa de 2015 (24 dias) e estação seca de 2016 (10 dias)); e zona polihalina (35 dias; estação seca de 2015 (5 dias), estação chuvosa de 2015 (22 dias) e estação seca de 2016 (8 dias)).

Figura 5. Densidade (ind.m<sup>2</sup>), Biomassa (g.m<sup>2</sup>), Riqueza, Diversidade H' e Pielou J' da macrofauna bentônica nas zonas euhalina e polihalina nas estações seca e chuvosa de 2015 e seca de 2016 no estuário Piraquê-Açú-Mirim.

Figura 6. Análise de ordenação (MDS) de assembleias macrobentônicas nas zonas euhalina e polihalina do estuário Piraquê-Açú-Mirim.

Figura 7. Modelos lineares baseados na distância (dbRDA), utilizando as propriedades sedimentares para explicar as mudanças das assembleias macrobentônicas no estuário do Piraquê-Açú-Mirim.

## SUMÁRIO

I.	Introdução geral	10
II.	Climatic effects in macrobenthic assemblages on an estuary in the Eastern Brazil Marine Ecoregion	13
	1. Introduction	13
	<ul> <li>2. Material and Methods</li> <li>2.1 Study area and sample design</li></ul>	15 15 16 16
	<ul> <li>3. Results</li></ul>	17 17 18 18 19 19 19 19 20
	4. Discussion	21
	5. References	24
	6. Table caption	31
	7. Figure caption	37
	8. Supplemental material documento .xlsx em separa	ıdo
III.	Conclusão	43
IV.	Referências	44
V.	Anexos	47

### 2 I. Introdução geral

3 As recentes alterações causadas pelas mudanças climáticas desencadearam respostas a nível global, alterando componentes biológicos (ex. mudanças na composição 4 e distribuição de plantas e animais), ambientais (ex. impactos em manguezais e recifes de 5 corais) e econômicos (ex. redução de produção agrícola e diminuição da quanlidade de 6 água (IPCC, 2001; Hughes et al., 2003). Os recentes acontecimentos são uma fração da 7 magnitude das mudanças previstas nos próximos séculos (IPCC, 2001). As mudanças 8 climáticas irão influenciar o ambiente marinho a partir do aumento da temperatura, 9 10 elevação do nível do mar, aumento na acidificação dos oceanos e intensificação dos eventos de inundações e secas (IPCC, 2001; Harley et al., 2006; Doney et al., 2012; Cook 11 et al., 2014). A elevação da temperatura média global proporciona mudança físicas e 12 químicas nos sistemas marinhos, como o derretimento do gelo polar, aumento da 13 temperatura da superfície do mar no ultimo século, mudanças na circulação atmosférica 14 e padrões de precipitação, como inundações e secas (IPCC, 2001; Harley et al., 2006; 15 Doney et al., 2012; Cook et al., 2014). Os ecossistemas, sua biodiversidade e 16 produtividade respondem de forma não-linear as mudanças climáticas (IPCC, 2001; 17 Harley et al., 2006; Hewitt and Thrush, 2009; Hoegh-Guldberg and Bruno, 2010). Essas 18 mudanças vão de grandes escalas, como uma seca em Solitary Islands Marine Park, 19 Austrália, onde alguns estuários intermitentes apresentaram condições hipersalinas 20 (Hastie and Smith, 2006), a local-específico, onde uma seca prolongada no estuário de 21 Santa Lucia, África do Sul, hipersalinizou algumas áreas, causando declínio da fauna 22 bêntica (Pillay and Perissinotto, 2008; MacKay et al., 2010). Os estuários apresentam 23 funções importantes e benefícios sociais, como a captura de sedimentos, a degradação 24 orgânica, oferecem refúgio e alimento à biota, sendo altamente produtivos, colocando os 25 estuários como ecossistemas costeiros de alta importância ecológica e socioeconômica, 26 especialmente diante das mudanças climáticas (Elliott and Whitfiel, 2011; Vilar et al., 27 2013; Worm et al., 2006). Apesar da importância dos serviços estuarinos, esses 28 29 ecossistemas são em grande parte influenciados por atividades humanas, como desenvolvimento costeiro, poluição, alteração de habitats e regimes hidrológicos (Edgar 30 et al., 2000; Kennish, 2002; Gusmão et al., 2016). A interação múltipla do estresse 31 humano pode afetar os ecossistemas através de um processo único, cumulativo ou 32 sinérgico (Vinebrooke et al., 2004; Hoegh-Guldberg and Bruno, 2010; Worm et al., 2006). 33 Exemplos de interações múltiplas são: i) efeitos nocivos dos metais pesados sobre as 34 assembleias bentônicas no estuário do Subaé (Brasil, Krull et al., 2014) e ii) menores 35 decréscimos durante estresses climático na produção e na biodiversidade em bancos de 36 Zostera noltii comparada a áreas eutrofizadas (antiga Z. noltii) no estuário do Mondego 37 (Dolbeth et al., 2007). 38

Os estuários são a zona de transição entre os rios e o mar, variando em forma, 39 tamanho, hidrografia, salinidade, características das marés, sedimentação e energia do 40 ecossistema entre regiões geográficas (Kennish, 2002; McLusky e Elliott, 2004). 41 Estuários apresentam gradientes de salinidade, variando de euhalina, com maior 42 influência marinha (30-40) a água doce, dominada pelos rios e drenagem continental (0-43 0,5, Sistema de Veneza 1958; McLusky e Elliott, 2004; Barros et al., 2012). Os 44 parâmetros físico-químicos, como salinidade, temperatura, oxigênio dissolvedo e 45 turbidez, variam de marés a estações (McLusky e Elliott, 2004, Dolbeth et al., 2011). 46 Alterações nos padrões de precipitação, como eventos e enchentes e secas, influenciam 47 os rios ondas e marés, que alteram os gradientes espaciais e temporais em curto e longo 48 prazo (Day et al., 2008; García-Rodriguez et al., 2013). Os eventos de inundações e secas 49 50 alteram a salinidade, a temperatura, a qualidade da água e a composição dos sedimentos nos estuários, que são chaves na distribuição da fauna bêntica (Attrill and Power, 2000; 51

#### 52 McLusky and Elliott, 2004).

Organismos bentônicos estuarinos estão distribuídos em diversos táxons, sendo 53 54 compostos principalmente por anelídeos, crustáceos e moluscos (Lee, 2008). A fauna pode ser classificada em diversos grupos funcionais como, depositívoros (superfície e 55 subsuperfície), carnívoros, omnívoros, filtradores e detritívoros (Jumars et al., 2015). As 56 57 assembleias bentônicas são amplamente utilizadas para estudos de monitoramento devido à sua importância para a estrutura e função do ecossistema, respondendo rapidamente a 58 mudanças naturais (ex. respostas diretas à salinidade e temperatura) e estresse 59 antropogênico (McLusky e Elliott, 2004; Pollack et al., 2011; Chen et al., 2011). Durante 60 eventos de seca, o aumento da salinidade e da temperatura do estuário podem afetar as 61 assembleias bentônicas por estresse fisiológico, alterando sua composição. 62 comportamento e ocorrência (Garcia-Rodriguez et al., 2013; Dittmann et al., 2015; Little 63 et al., 2016). Estas alterações podem afetar as funções dos ecossistemas e os processos 64 biogeoquímicos de sistemas bentônicos em pequenas e grandes escalas, como i) bancos 65 de Zostera noltii no estuário do Mondego (Portugal), que diminuem a produção e a 66 biodiversidade (Dolbeth et al., 2007), ii) decréscimo na produção de espécies 67 comercialmente importantes devido à variabilidade climática no estuário do Mondego 68 (Portugal; Grilo et al., 2011), iii) Alta variabilidade na salinidade diminuem a abundância 69 70 e diversidade de bivalves e crustáceos, bem como a ocorrência de espécies bentônicas não registadas no sistema em densidades elevadas (Haminoea natalensis e a espécie 71 exótica Tarebia granifera) que podem causar sérias ameaças ao estuário de Santa Lucia 72 73 (África do Sul; Pillay e Perissinotto, 2008) e iv) reduzir a produção de plâncton com efeitos em cascata em níveis tróficos superiores no estuário de Neuse (Estados Unidos; 74 Wetz et al., 2011). Eventos de seca atuam de forma diferente entre estuários e zonas 75 estuarinas, podendo causar efeitos deletérios na carga de nutrientes, taxa de oxigênio 76 dissolvido e produção primária, além de afetar a composição de espécies devido a 77 presença de espécies marinhas (IPCC, 2001; Vinebrooke et al., 2004; Harley et al., 2006; 78 79 Hoegh-Guldberg and Bruno, 2010; MacKay et al., 2010; Dolbeth et al., 2011; Elliott and 80 Whitfield, 2011; Mcleod et al., 2011; Montagna and Palmer, 2012; Wetz and Yoskowitz, 2013). 81

Entender os efeitos das secas e inundações nos estuários é de extrema importância 82 para compreender os potenciais impactos climáticos nesses ecossistemas (Bernardino et 83 al., 2015), bem como melhorar a gestão dos ecossistemas no próximo século (Dai, 2011, 84 2013; Cook et al. 2014). A avaliação a longo prazo da oscilação natural permite esclarecer 85 os padrões, sendo extremamente útil para a gestão, fornecendo modelos mais precisos 86 (Dolbeth et al., 2011, Mcleod et al., 2011; Bernardino et al., 2016). Surpreendentemente, 87 a literatura sobre eventos de seca em estuários é escassa numa perspectiva global, apenas 88 16 estuários possuem monitoramento a longo prazo. Mais da metade dos estuários 89 monitorados estão localizados na Austrália (5) e nos Estados Unidos (4), com dados 90 publicados principalmente durante o século 21 (Attrill e Power, 2000, Pillay e 91 Perissinotto, 2008; Dai, 2011, 2013). Dez estuários têm um grande acompanhamento 92 temporal (> 10 anos), com informações bem documentadas e grande conjunto de dados 93 sobre eventos de seca em assembleias bentônicas estuarinas, sendo estes, i) Nueces, 94 95 Lavaca-Colorado, Guadalupe and Caloosahatchee nos Estados Unidos (4), ii) Europa (3, Golfo do Leão na França, estuários do Mondego em Portugal, e do Guadalquivir na 96 Espanha), iii) Australia na foz Murray e lagoas Coorong no sistema Murray-Darling, e 97 iv) estuário da Santa Lucia na Africa do Sul (Montagna e Palmer, 2012; Palmer et al., 98 2015). O estuário de Santa Lucia funciona como um modelo vivo das previsões sobre 99 intensificações de eventos de seca durante o século XXI, este sistema apresenta longos 100 períodos naturais de seca, ex. cinco anos, que acarretam em fechamento da foz e 101

condições de hipersalidade, evidenciando a necessidade de compreender estes eventos
(Pillay e Perissinotto, 2008; MacKay et al., 2010). Em geral, faltam informações sobre o
monitoramento de estuários na América do Sul, África e Ásia, que destacam as maiores
ameaças a esses ecossistemas.

As mudanças climáticas em escalas espaciais e temporais mudarão os padrões de 106 107 precipitação, salinidade e temperatura (Attrill e Power, 2000; Dai, 2011, 2013). Estas alterações influenciarão os padrões macrofaunais e as funções do ecossistema de forma 108 diferente entre áreas preservadas e impactadas (por exemplo, banco de Zostera noltii e 109 área eutrofizadas no estuário do Mondego; Dolbeth et al., 2007). A compreensão dos 110 efeitos dos estressores naturais e antropogênicos é essencial para uma melhor gestão dos 111 estuários (Vinebrooke et al., 2004; Sweetman et al., 2010; Dolbeth et al., 2011; 112 Bernardino et al., 2015). O estuário Piraque-Açú-Mirim, localizado na Ecorregião 113 Marinha do Leste Brasileiro em uma área relativamente bem preservada, foi utilizado 114 como um estudo de caso para a compreensão dos efeitos climáticos sobre as assembleias 115 bentônicas. O monitoramento do estuário do Piraque-Açú-Mirim é um dos mais extensos 116 dos estuários brasileiros (n = 14 meses, Bernardino et al., 2016) sendo vantajoso, pois o 117 estuário está em uma área com evidência de significativo aumento na temperatura média 118 anual durante a última década (Bernardino et al., 2015). Aqui apresentamos um 119 monitoramento temporal (tempo de meses) de precipitação, salinidade e temperatura da 120 água, e sedimentos na região estuarina, cujo objetivo é i) identificar mudanças temporais 121 nos sedimentos e conjuntos bentônicos no estuário Piraque-Açú-Mirim; e ii) relacionar a 122 123 influência potencial do evento de seca às propriedades sedimentares e da macrofauna bentônica. Nossa hipótese é que as assembleias macrobentônicas seriam impactadas por 124 fatores ambientais com correlação temporal a mudanças na salinidade e temperatura da 125 água. 126

# II. Climatic effects in macrobenthic assemblages on an estuary in the Eastern Brazil Marine Ecoregion

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### 138 Abstract

In estuaries, changes in rainfall, including flood and drought events may influence water 139 salinity, temperature and sediment grain size, with consequences to the composition of 140 benthic macrofaunal assemblages. With predicted climate change impacts on coastal 141 ecosystems worldwide, understanding the effects of natural stressors is essential to 142 evaluate effects on the biota and on ecosystem function of estuarine ecosystems. This 14-143 month study on an estuary in the Eastern Brazil Marine Ecoregion, revealed temporal 144 changes in sediment composition and macrofaunal benthic assemblages in response to a 145 drought event likely associated with the 2015-2016 El-Niño. We detected seasonal 146 changes on sedimentary properties, with a predominance of clay particles and low 147 sedimentary Chlorophyll-a, following the drought event. Macrofaunal assemblages in 148 euhaline and polyhaline zones were similar, but observed changes ocurred during the dry 149 season in 2016. Water temperature influenced macrofaunal diversity and species 150 composition within the estuary, supporting the occurrence of assemblages shifts during 151 drought events as well as a decrease in diversity and abundance of dominant taxa in the 152 Piraquê-Açú-Mirim estuary. Long-term monitoring stations will be invaluable in future 153 154 years of climate change in order to better quantify impacts on estuaries and other coastal 155 ecosystems with the likely increase in intensity and frequency of droughts in some regions of Brazil. 156

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158 Keywords: Rainfall decrease, Monitoring, Benthic ecology, Climate change

## 159160 **1. Introduction**

Climate change and the intensification of floods and droughts events are expected 161 to significantly alter marine ecosystems (IPCC, 2001; Harley et al., 2006; Dai, 2011, 162 2013; Doney et al., 2012; Cook et al., 2014). Rise in global mean temperature provides 163 physical and chemical change in marine systems, such as polar ice melting, rising sea 164 surface temperatures in the last century, changes in atmospheric circulation, and 165 precipitation patterns such as floods and droughts (IPCC, 2001; Harley et al., 2006; Dai, 166 2011, 2013Doney et al., 2012; Cook et al., 2014). Climate changes promove non-linear 167 changes on ecosystems function, biodiversity and productivity due multiple stressors 168 interaction (IPCC, 2001; Harley et al., 2006; Hoegh-Guldberg and Bruno, 2010). These 169 changes range from large scales, such as a drought in Solitary Islands Marine Park, 170 Australia, where some intermittent estuaries have hypersaline conditions (Hastie and 171 Smith, 2006), to site-specific, as a prolonged drought in the Santa Lucia estuary, South 172 Africa (Pillay and Perissinotto, 2008; MacKay et al., 2010). Estuaries present important 173 functions and societal benefits, such as sediment trapping, organic degradation, offer 174 refuge and food to the biota, being highly productive, these services place estuaries as 175 176 coastal ecosystems of high ecological and socioeconomic importance, especially in face of climatic changes (Elliott and Whitfiel, 2011; Vilar et al., 2013; Worm et al., 2006). 177

Despite the importance of estuarine services, these ecosystems are largely influenced by 178 human activities, such as coastal development, pollution and altered habitat (Edgar et al., 179 180 2000; Kennish, 2002; Gusmão et al., 2016). The multiple interaction of human stress can impact ecosystems through single, cumulative or synergistic process (Vinebrooke et al., 181 2004; Hoegh-Guldberg and Bruno, 2010; Worm et al., 2006), as i) Harmful effects of 182 183 heavy metals on benthic assemblages in the Subaé estuary (Brazil; Krull et al. 2014) and ii) lower decreases during climatic stresses in production and biodiversity in Zostera noltii 184 banks compared to the eutrophic area in the Mondego estuary (Dolbeth et al., 2007). 185

Estuaries are the transitional zone between rivers and the sea, which vary widely 186 in hydrological, morphological and chemical conditions. Physicochemical parameters, as 187 salinity, temperature, dissolved oxygen and turbidity, vary from tides to seasons. Haline 188 gradients are evidentes across the estuary, from euhaline (30-40) to freshwater (0-0.5, 189 Venice System 1958; Kennish, 2002; McLusky and Elliott, 2004; Dolbeth et al., 2011). 190 Changes in rainfall patterns, as floods and droughts events, influences riverine, wave or 191 tidal processes, which changes spatial and temporal gradients by short to long-terms (Day 192 et al., 2008; Garcia-Rodriguez et al., 2013). Floods and droughts events change salinity, 193 temperature, water quality and sediment composition in estuaries, which are keys of 194 benthic assemblages distribution and occurrence (Attrill and Power, 2000; McLusky and 195 Elliott, 2004). 196

197 Changes from droughts seasons will likely promote a large number of responses, from i) maintance or increased in primary production (Molen and Perissinotto, 2011), ii) 198 changes in sedimentation rates and its biogeochemical interactions (Dittmann et al., 2015; 199 Montagna and Palmer, 2012) to iii) hipersalinization (Pillay and Perissinotto, 2009; 200 Palmer et al., 2015), iv) close of the estuarine mouth (Pillay and Perissinotto, 2009; 201 MacKay et al., 2010). Changes on estuarine ecosystems by droughts can be intensifyed 202 or masked due human pressure as pollution, proximity to large cities, deforestation, 203 agriculture, dams, and others (Kennish 2002; Dolbeth et al., 2007; Cardoso et al., 2008). 204 Again, the responses of macrofaunal assemblages to the drought will be in single, 205 cumulative or synergistic processes, ranging from i) increase in density, diversity, number 206 and biomass of taxa, ii) do not respond imediatly to the drought by external influece, as 207 support of nutrientes from non-treated waste, or iii) presents the complete inverse, as 208 decrease in density, diversity, number and biomass of taxa (IPCC, 2001; Pillay and 209 Perissinotto, 2008, 2009; Grilo et al., 2009, 2011; Table 5, references therein). 210

Benthic assemblages are widely used to monitoring studies due its importance to 211 the structure and function of the ecosystem, responding rapidly to natural changes (e.g. 212 direct responses to salinity and temperature) and anthropogenic stress (McLusky and 213 Elliott, 2004; Kotta et al., 2009; Pollack et al., 2011). During drought, the increase of 214 estuarine salinity and temperature can influence benthic assemblages by physiological 215 stress changing its composition, behavior and occurrence (Garcia-Rodriguez et al., 2013; 216 Dittmann et al., 2015; Little et al., 2016). These changes may alter ecosystem functions 217 and biogeochemical processes provided by benthic assemblages on small and large scales, 218 as i) decrease of commercially important species production due climate variability in the 219 Mondego estuary (Portugal; Grilo et al., 2011), ii) high variability changes in salinity 220 221 decrease abundance and diversity of bivalves and crustaceans, as well unrecorded macrofaunal species to the system in high densitys (Haminoea natalensis and the alien 222 species Tarebia granifera) which can cause serious threats to the Santa Lucia Estuary 223 (South Africa; Pillay and Perissinotto, 2008) and iii) reduce plankton production with 224 cascade effects on higher trophic levels in Neuse estuary (United States; Wetz et al., 225 2011). 226

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Understand the effects of droughts and floods on estuaries is of utmost importance

to understand potential climate impacts on these ecosystems (Bernardino et al., 2015), as 228 well improve ecosystem management in the next century (Dai, 2011, 2013; Cook et al., 229 2014). Long-term evaluation of natural oscillation allows elucidating patterns, being 230 extremely useful for management by more accurate models (Dolbeth et al., 2011; Elliott 231 and Whitfield, 2011; Mcleod et al., 2011; Bernardino et al., 2016). Drought literature' on 232 233 estuaries is scarce from a global perspective, only 16 estuaries have been surveyed in view to analyse drought effects, most of those as a result of long term monitoring. Ten 234 estuaries have large temporal monitoring (>10 years), with well documented information 235 to droughs events on benthic assemblages, with data mostly published during the 21 236 century (Attrill and Power, 2000; Montagna et al., 2002; Pillay and Perissinotto, 2008; 237 Dai, 2011, 2013; Palmer et al., 2015; Table 5, references therein). Santa Lucia estuary 238 works as a living model of possible drought intensification during the 21 century, by 239 natural several seasons of droughts, as five years of drought, mouth closure and 240 hypersaline conditions, evidencing the necessity to better understand the drought events 241 (Pillay and Perissinotto, 2008, 2009; MacKay et al., 2010). In general, are lacking 242 information of monitoring estuaries in South America, Africa and Asia, which highlight 243 majors' threats to these ecosystems. 244

Climate change at spatial and temporal scales will change patterns of precipitation, 245 246 salinity and temperature (Attrill and Power, 2000; Dai, 2011, 2013). These changes will influence macrofaunal patterns and ecosystem functions differently from preserved and 247 impacted areas (e.g. Zostera noltii bed and eutrophic area in the Mondego estuary, 248 Portugal; Dolbeth et al., 2007). Understanding the effects of natural and anthropogenic 249 stressors is essential to better manage estuaries (Vinebrooke et al., 2004; Sweetman et al., 250 2010; Dolbeth et al., 2011; Bernardino et al., 2015). The Piraque-Açú-Mirim estuary, 251 located in the East Brazilian Marine Ecoregion in a relatively well preserved area, was 252 used as a case study for the understanding of climatic effects on benthic assemblages. The 253 monitoring of the Piraçê Açú estuary is one of the most extensive of Brazilian estuaries 254 (n=14 months; Bernardino et al., 2016) and advantageous because the estuary is in an 255 area with evidence of significant decrease in annual average rainfall over the last decade 256 (Bernardino et al., 2015). Here we present a temporal (time scale of months) monitoring 257 of precipitation, salinity and temperature of water, and sediments in the estuarine region, 258 whose objective is i) identify temporal changes in sediments and benthic assemblages in 259 the Piraquê-Açú-Mirim estuary; and ii) relate the potential influence of the drought event 260 to the sedimentary and macrofaunal properties. Our hypothesis is that macrofaunal 261 assemblages would be different over time and between haline zones responding to 262 changes in salinity and water temperature. 263

## 265 **2. Material and Methods**

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### 266 2.1. Study area and sample design

This study was carried at the Piraquê-Açú-Mirim estuary (17°58'S; 40°00'W), 267 within the Eastern Brazil Marine Ecoregion (Bernardino et al., 2015; Fig. 1). This estuary 268 has a Y-shape morphology with extensive mangroves, about 12 Km<sup>2</sup>, composed of 269 Rhizophora mangle, Laguncularia racemosa and Avicennia schaueriana; and tidal flats, 270 271 which are part of a Sustainable Development Reserve. The estuary is under a semi-diurnal microtidal regime (< 2m). Coastal development led to mangrove removal in some areas 272 on the estuarine margin for agriculture, house construction and piers to fishermen access 273 the river (Bernardino et al., in review). 274

The Piraquê-Açú-Mirim estuary has an euhaline (30-40) and a polyhaline (18 to 30) zones, which were defined according to previous environmental data (Bernardino, unpublished) and from pilot studies taken in April and May 2015 with a

conductivity/temperaute data-logger (OnSet) and a multiparameter (Venice System, 278 1958; Gimenez et al., 2005; Mariano and Barros, 2015; Figure 1). We selected two 279 280 random study sites at each haline zone, one at each river margin distant about 100 m (Euhaline: Site 1 19°56'17.1" S, 40°10'37" W; Site 2 19°56'38.5" S, 40°10'26.1" W; 281 Polyhaline Site 1 19°55'09.5" S, 40°12'28.5" W; Site 2 19°56'18"S, 40°12'24.7" W). Each 282 283 site was subdivided in two sampling areas distant 10 m from each other on the intertidal mudflat region, distant at least 1m from the nearby mangrove forests. Sediments and 284 macrofaunal assemblages were sampled monthly from May 2015 to June 2016 at the same 285 areas. At each area, three faunal samples and a sediment sample (three replicates, 286 0.008m<sup>2</sup>) were collected for analysis of Particle size, Total organic content and Calcium 287 carbonate. A superficial sediment sample (15g, 0-3 cm) was sampled for Chlorophyll-a 288 and Phaeopigments analysis (Quintana et al., 2015). Samples were grouped by seasons, 289 Dry 2015 = May to September 2015, Wet 2015 = October 2015 to March 2016, and Dry 290 2016 = April to June 2016 (Reboita et al., 2010). 291

Mean daily rainfall were obtained from the website of the Brazilian National 292 Water Agency (ANA, 2016; Rainfall stations: 1940002, 1940021, 1940039, 194,0042, 293 2549007 and 2549087; Years from 1948 to 2016) and from the local climatic station at 294 UFES (2015-2016). Water pH, total dissolved solids (TDS, ppt) and dissolved oxygen 295 296 (mg/L, DO) were sampled at each area within the euhaline and polyhaline zones prior to sediment sampling using a HANNA multiparameter near the mud flat (1 m depth, low 297 tide). The monitoring of salinity and Temperature (°C) were carried using an OnSet 298 299 datalogger (U24-002-C) deployed continually, during seasons of 5 to 24 days at 1 meter above the bottom (~3 meters depth, low tide). A total of 40 days were monitored at the 300 euhaline zone (Dr 2015 (6d), Wet 2015 (24d) and Dry 2016 (10d)); and 35 days at the 301 302 Polyhaline zone (Dry 2015 (5d), Wet 2015 (22d) and Dry 2016 (8d)). Except during the Dry 2015 season where only one spring tide was monitored, all other seasons were 303 monitored during spring and neap tides. 304

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## 306 2.2. Laboratory analysis

Macrofaunal samples were sieved in the field (500µm mesh size) using local water 307 and preserved in 70% Ethanol until analysis. In the laboratory, all organisms were sorted 308 and identified to the lowest taxonomic level. After sorting, total macrofaunal biomass 309 (wet weight) was determined using a 0.0001g precision balance. The sediment particle 310 size was determined by dry-sieving and reported as mud sediment fraction (i.e. particles 311 sizes >2, 2-1,1-0.5, 0.5-0.25, 0.25-0.125, 0.125-0.063 and <0.063 mm). Total organic 312 content was measured by loss on ignition over 4 h at 500°C, after drying the samples at 313 60°C for 48 h. Calcium carbonate (CaCO3) contents were determined by adding chloric 314 acid (10%; 2 mL) till effervescence ceased and then dried at 60°C for 48 h. Chlorophyll-315 a and phaeopigment were analyzed according to Lorenzen (1967). In laboratory, 316 chlorophyll-a concentrations were measured spectrophotometrically (absorbance read at 317 430 and 665 nm) after extraction with 100% acetone. Phaeopigments were determined 318 after acidification with 0.1 N HCl (Lorenzen, 1967). 319

320

### 321 2.3. Statistical analysis

Total daily rainfall at the Piraquê-Açú-Mirim region (rainfall stations of ANA local and climatic station at UFES) were grouped at scales of months, seasons and years (mean and standard deviation) to describe the drought event. Water salinity and temperature were monitored by conductivity and temperature data-loggers at euhaline and polyhaline zones across the three seasons during the study (Dry 2015, Wet 2015, Dry 2016), and spanned major tidal changes (spring x neap) at each haline zone. PSS-78 was used to calculated salinity using the logger conductivity and temperature measures(UNESCO, 1981).

Based on the unbalanced design, a one-way Analysis of Variance (ANOVA) model was used to compare salinity and temperature accross tides and seasons. Sedimentary properties (Mud content, Total organic matter, Calcium carbonate, Chlorophyll-*a* and Phaeopigment) and macrofaunal assemblages (Density, Biomass, Species richness, Shannon H' and Pielou J') were compared between haline zones (euhaline and polyhaline) and seasons (Dry 2015, Wet 2015, Dry 2016). If significant, *a post hoc* pairwise comparison Tukey-Kramer test was used, due unequal sample sizes (Dunnett, 1980).

Spatial and temporal changes of the benthic assemblages were assessed by density 337 (ind.m<sup>2</sup>), biomass (g.m<sup>2</sup>), taxa composition and diversity analysis (Species richness, 338 Shannon H 'and Pielou J'). Analysis of Multidimensional Similarity (MDS) was applied 339 using transformed square root abundance of all taxa in a similarity matrix (Bray-Curtis) 340 to evaluate changes between seasons, and euhaline and polyhaline zones. The significance 341 of clustering of MDS was assessed using the SIMPROF test and significance tests were 342 determined by ANOSIM (Warwick and Clarke 1993). Changes in benthic assemblages 343 and sedimentary parameters were analyzed using PERMANOVA (permutational analysis 344 of variance; Anderson et al., 2008). Two factors were used in PERMANOVA, seasons 345 (Dry 2015, Wet 2015 and Dry 2016), and haline zones (euhaline and polyhaline). 346 PERMANOVA of benthic assemblages was conducted using a Bray-Curtis similarity 347 matrix, transformed by the fourth root. The sedimentary parameters were transformed by 348 Euclidean distance. 349

Temporal correlations between water salinity and temperature with sedimentary 350 parameters (Mud content, Total organic matter, Calcium carbonate, Chlorophyll-a and 351 Phaeopigment), and macrofaunal assemblages properties (Density, Biomass, Species 352 richness, Shannon H 'and Pielou J') and composition were tested by Multiple linear 353 regression analysis. Valid regression models were tested by an ANOVA (global quality of 354 fit; Legendre and Legendre, 2012). Water salinity and temperature were chosen as 355 predictors of pluviometric and macrofaunal changes, due to their inverse relationship with 356 rainfall variations and direct effects on macrofaunal assemblages (Palmer et al., 2015). 357

DistLM (distance-based linear model) routines (McArdle and Anderson, 2001; 358 Anderson et al., 2008) were performed Selection = step-wise, selection criterion = 359 adjusted AICc); and applied to sedimentary parameters and macrobenthic assemblages 360 (Clarke and Gorley, 2006; Anderson et al., 2008). Benthic assemblage data were analyzed 361 using the Bray-Curtis similarity matrix, transformed by the fourth root. Sedimentary 362 parameters were transformed by Euclidean distance. BIO-ENV procedure was applied to 363 relate multivariate patterns of macrobenthic assemblages by sedimentary parameters 364 (classification of Spearman, *p* between the two matrices of similarity). 365

Analyzes of Diversity, MDS, SIMPROF, ANOSIM, PERMANOVA, RELATE, DISTLM and BIO-ENV were performed using the software PRIMER v 6.0 with the PERMANOVA + add-on package (Clarke and Gorley, 2006; Anderson et al., 2008). ANOVA with *post hoc* Tukey-Kramer test and Multiple linear regression analyzes were performed Using SPSS v 20.0 software (IBM SPSS Statistics Inc., Chicago, IL, USA).

371372 **3. Results** 

373 *3.1. Drought characterization* 

374

Rainfall varied drastically during the study seasons in the Piraquê-Açú-Mirim estuary, with higher monthly volumes during the Dry season of 2015 (62.5±52.5 mm/mo), which were similar to the mean historic rainfall for the season (Dry 1948-2014, 62.6±19.5

mm/mo; Figure 2). However, during the Wet season of 2015 (46.9±29.7 mm/mo) and the 378 following the Dry season of 2016, (13.3±5.9 mm/mo), rainfall was lower than historical 379 means, with strong reductions of 40 to 80% in rainfall volume. The lowest montly rainfall 380 volume occurred during the Dry season of 2016, with values of 8.9 to 11 mm/mo in 381 February, April and May 2016 (Figure 2). Historical data indicated that the high rainfall 382 383 at the Piraquê-Açú-Mirim estuary occurred in December 2013, with a mean monthly rainfall of 649.8±115.5 mm. Our data revealed that during the last decade, the Piraquê-384 Açú-Mirim estuary has been under a continuous decline in mean rainfall, with the lowest 385 rainfall averages of the last 50 years occurring in 2015 and 2016 (Figure 2). 386

As expected, the mean water salinity of euhaline and polyhaline zones were 387 different at all seasons (dry and wet seasons of 2015 and Dry season of 2016), and 388 between spring and neap tides. Spring tide salinity was always higher in the euhaline zone 389 if compared to the polyhaline zone (26.7 to 26.9 and 23.5 to 24.1, respectively; ANOVA 390 p < 0.001 Figure 3). The same occurred with neap tide salinity ranges, which were higher 391 in the euhaline zone during all seasons sampled (Figure 3, ANOVA p < 0.001). However, 392 tidal patterns in salinity ranges changed within each zone across seasons. In the Euhaline 393 zone, salinity range during spring tides was significantly different to neap tides during the 394 Wet season of 2015 (ANOVA F 37.414, p 0.000; Figure 3), not during the Dry season of 395 2016. In the Polyhaline zone, salinity ranges between spring and neap tides was 396 significantly different during the Wet season of 2015 (ANOVA F 26.403, p 0.000) and the 397 Dry season of 2016 (ANOVA F 6.610, p 0.011). However, salinity ranges at both spring 398 and neap tides in the Polyhaline zone increased significantly, with lower oscilation, 399 between the Dry and Wet seasons of 2015 to Dry season of 2016 (ANOVA F 24.351, p 400 0.000; Tukey-Kramer p 0.000, both seasons; Figure 3). 401

402 Spring tide temperatures in euhaline zones were constantly higher if compared to the polyhaline zones in Piraquê-Açú-Mirim estuary (ANOVA p<0.000; Figure 4). During 403 the Wet season of 2015 and Dry season of 2016, neap tide temperatures in the euhaline 404 zone were lower than polyhaline (ANOVA p<0.000 Figure 4). Tidal patterns in 405 406 temperature ranges changed within each zone across seasons. In the Euhaline zone, temperature range during spring tides was significantly different to neap tides during the 407 Wet season of 2015 (ANOVA F 6.938, p 0.009) and Dry season of 2016 (ANOVA F 408 71.762, p 0.000). In the Polyhaline zone, temperature ranges between spring and neap 409 tides was significantly different during the Wet season of 2015 (ANOVA F 4.009, p 410 0.047), not during the Dry season of 2016 (Figure 4). However, temperature ranges at 411 both spring and neap tides in the euhaline zone increased significantly, with lower 412 oscilation, between Dry season of 2015, Wet season of 2015 and Dry season of 2016 413 (ANOVA F 298. 960, p 0.000; Tukey-Kramer p 0.000). The same pattern was evidenced 414 in the polyhaline zone temperature, which increased significantly, with lower oscilation 415 between Dry season of 2015, Wet season of 2015 and Dry season of 2016 (ANOVA F 416 572.984, p 0.000; Tukey-Kramer p 0.000 all seasons; Figure 4). 417

- 418
- 419 *3.2. Spatial changes between haline zones*

420 *3.2.1. Water column and sediment* 

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Water column pH, Total dissolved solids (TDS) and Dissolved oxygen (DO) were similar at the euhaline and polyhaline haline zones (Table 1). All sediment properties (Mud content, Total organic content, Calcium carbonate, Chlorophyll-*a* and Pheopigments) were also similar between the euhaline and polyhaline zones (PERMANOVA, F 0.467, *p* 0.79; Table S1).

#### 428 *3.2.2. Macrofaunal assemblages*

429 A total of 3,736 individuals distributed in 32 taxa were sampled during the study 430 period. Annelida (17 taxa), Crustacean (7 taxa) and Mollusca (6 taxa) were predominant. 431 Macrofaunal density, biomass, richness, Shannon H' log e and Pielou J' were similar at 432 the euhaline and polyhaline zones (ANOVA, p > 0.05; Table S2).

433 Annelida was the most abundant taxa in the euhaline and the polyhaline zones. The polychaetes Magelonidae, Sternaspidae, Capitellidae, Goniadidade and Paraonidae 434 euhaline (234,000±51,282 ind.m<sup>2</sup>, 89.3%) dominated the and polvhaline 435 (179,375±39,449 ind.m<sup>2</sup>, 87.5%) zones of Piraquê-Açú-Mirim estuary. Macrofaunal 436 composition markedly changed between euhaline and polyhaline zones. The crustacean 437 Cirolanidae  $(375\pm29 \text{ ind.m}^2, 0.1\%)$  only occured at the euhaline zone, while the 438 polychaetes Amphinomidae (125±9.6 ind.m<sup>2</sup>, 0.1%) and Sabellariidae (125±9.6 ind.m<sup>2</sup>, 439 0.1%), and the crustaceans Atyidae ( $125\pm9.6$  ind.m<sup>2</sup>, 0.1%), Alpheidae ( $125\pm9.6$  ind.m<sup>2</sup>, 440 0.1%) and Cumacea ( $125\pm9.6$  ind.m<sup>2</sup>, 0.1%), the gastropod Vitrinellidae ( $750\pm49.1$ 441 ind.m<sup>2</sup>, 0.4%) and Sipuncula (250±13.6 ind.m<sup>2</sup>, 0.1%) only occured at the polyhaline zone 442 of Piraquê-Açú-Mirim estuary. About 25% (8 taxa) of the sampled taxa were rare and 443 represented less than 1% of the relative abundance (Table 2; Table S3). 444

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446 *3.3. Temporal changes in the estuary* 

447 3.3.1. Water column and sediment

448

449 Water column pH, Total dissolved solids (TDS) and Dissolved oxygen (DO) were similar at Dry and Wet seasons of 2015 and Dry season of 2016 at both haline zones 450 (Table 1). Dissolved Oxigen was lower at the euhaline and polyhaline zones during the 451 452 Dry season of 2016 (Table 1). The sediment Mud content increased significantly from Dry 2015 and Wet 2015 to Dry 2016 (ANOVA, F 10.023, p 0.001). Sediment Chlorophyll-453 a was also significantly lower during the Dry season of 2016, if compared to the two 454 previous seasons (ANOVA, F 4.71, p 0.01; Table S1). Sediment Pheopigments decrease 455 significantly between Dry season of 2015 and Wet season of 2015 (ANOVA p 0.033). 456 Sedimentary Mud content, Chlorophyll-a and Pheopigments of euhaline and polyhaline 457 zones in Piraquê-Açú-Mirin estuary vary significantly between seasons (PERMANOVA, 458 F 11.295, p 0.001; Table S1). 459

Temporal changes in sediment Chlorophyll-a and Pheopigments were influenced 460 by salinity in euhaline and polyhaline zones, respectively (Multiple linear regression F 461 462 2.520; p 0.027; F -3.060; p 0.01). Changes in Mud content (Multiple linear regression F 2.394; p 0.034) was influenced by temperature in euhaline zone. Changes in Mud content 463 (Multiple linear regression F 4.034; p 0.002) and Pheopigments (Multiple linear 464 regression F -2.202; p 0.05) were influenced by temperature in polyhaline zone. 465 Sedimentary properties (Mud content, Total organic content, Calcium carbonate, 466 Chlorophyll-a and Pheopigments) in euhaline and polyhaline zones were similar with 467 different changes across the Dry 2015, Wet 2015 and Dry 2016 seasons in Piraquê-Açú-468 Mirim estuary (PERMANOVA, F 1.1101 p 0.336). 469

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#### 471 *3.3.2. Macrofaunal assemblages*

472 Seasonal changes on macrofaunal density decreased significantly between Dry 473 season of 2015 than Wet season of 2015 (p 0.007), and with Dry season of 2016 (p 0.002) 474 in both haline zones (Figure 5). Macrofaunal richness decreased significantly between 475 Dry season of 2015 than Wet season of 2015 (p 0.000), and to Dry season of 2016 (p476 0.001) in both haline zones. Macrofaunal diversity Shannon decreased significantly from 477 the Dry to Wet seasons of 2015 (p 0.028), whereas macrofaunal evenness (J') increase significantly in Dry season of 2015 and 2016 (*p* 0.008). Macrofaunal assemblages of
euhaline and polyhaline zones vary significantly between seasons in Piraquê-Açú-Mirin
estuary (PERMANOVA, F 11.295, *p* 0.001; Figure 5, Table S2).

Macrofaunal composition markedly changed across seasons in the Piraquê-Açú-481 Mirim estuary. During the Dry and Wet season of 2015, Magelonidae, Sternaspidae, 482 483 Capitellidae, Goniadidade and Paraonidae dominated the euhaline (91%) and polyhaline (87.8-90%) zones. However, the top five ranked taxa changed during the Dry season of 484 2016, with Oligochaete becoming a dominant taxa, with a marked decrease in Paraonidae 485 at the euhaline and polyhaline zones (Table 2). In overall, during the Dry season of 2016 486 there was a decrease in the dominance of the top ranked macrofaunal taxa (<84%), if 487 compared to the two previous seasons Dry 2015 (>87%) and Wet 2015 (~90%). 488

Sixteen macrofaunal taxa only occured in a specific season, and most of then had low relative abundances. Eight taxa (25% of total taxa) only occurred during the Dry season of 2015, while the mussel Vitrinellidae only occurred in the Wet season of 2015, all in low relative abundances (<0.1%). Seven taxa only occurred during the Dry season of 2016, six in low relative abundances, with Oligochaete in high relative abundance to euhaline zones (5.9%) and polyhaline zones (7.8%) to both haline zones, being a top five ranked taxa in Piraquê-Açú-Mirin estuary (Table 2).

Water temperature influence the increase in Macrofaunal Pielou J' in euhaline zones (Multiple linear regression F 3.217, p 0.007), and the decrease in Macrofaunal Richness in polyhaline zones (Multiple linear regression F -3.889, p 0.002). Macrofaunal assemblages in euhaline and polyhaline zones were similar (dominant taxa) across the Dry 2015, Wet 2015 and Dry 2016 seasons in Piraquê-Açú-Mirim estuary (PERMANOVA, F 1.0386, p 0.406).

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#### 503 3.3.3. Multivariate analysis

Multivariate analysis revealed a marked seasonal change in benthic assemblages 504 in the Piraquê-Açú-Mirim estuary (ANOSIM, R 0.305, p 0.001). Macrofaunal 505 assemblages differed between seasons (Dry 2015, Wet 2015 and Dry 2016; 506 PERMANOVA, Pseudo-F 11.693, p 0.001). The dissimilarity was mainly influenced by 507 13 of the 32 macrofaunal taxa (Global Test, R 0.951, p 0.1%). Annelids mostly influenced 508 the observed seasonal changes with the 10 top ranked taxa (Figure 6; Table 2). 509 Magelonidae, Spionidae, Sternaspidae, Nereididae and Goniadidae decreased in density 510 across the seasons in euhaline zones and polyhaline zones. Paraonidae decreased in 511 density from the Dry season of 2015 to the Wet season of 2015 and were absent in the 512 Dry season of 2016 in both zones. Oligochaete only occured in the Dry season of 2016 in 513 euhaline and polyhaline zones of Piraquê-Açú-Mirim estuary (Figure 6; Table 2). 514

Macrofaunal assemblages in the euhaline zones and polyhaline zones were 515 dominated by the polychaetes Magelonidae, Sternaspidae, Capitellidae and Goniadidae 516 (46.3%; SIMPER analysis). Macrofaunal composition were influenced by the decrease in 517 density of the top dominant taxa (Magelonidae, Sternaspidae, Capitellidae, Goniadidae) 518 and less representative taxa (Spionidae, Nereididae, Phyllodocidae, Scalibregmatidae and 519 Nemertea), as well the increase of Pilargidae and Cossuridae across Dry season of 2015 520 521 to Dry season of 2016. Change in macrofaunal assemblage composition between seasons were influenced by the absence of Paraonidae and Tellinidae in Dry season of 2016 522 compared to Dry and Wet seasons of 2015 of Piraquê-Açú-Mirim estuary (Figure 6; Table 523 2). 524

525 Sedimentary mud content contributed to the seasonal changes on macrofaunal 526 assemblages (DISTLM, Pseudo-F 13.392, *p* 0.001; Table 3). Changes in mud content 527 explained 12.7% of macrofaunal assemblage variation (BEST R 0.107; Table 4, Figure 528 7).

Water temperature influence significantly the decrease in macrofaunal density in euhaline zones and polyhaline zones. Water temperature in the euhaline zone influence the decrease in density of Phyllodocidae and Cirratulidae (Multiple linear regression ANOVA F 14.500, p 0.001), whereas in the polyhaline zones influence the decrease in density of Paraonidae, Lumbrineridae, Goniadidae and Phyllodocidae (Multiple linear regression ANOVA F 45.484, p 0.000). The polychaete Phyllodocidae was the unique taxa influenced by temperature in euhaline zones and polyhaline zones.

- 536537 4. Discussion
- 538

Changes in rainfall patterns, with increases in flood and drought events, are 539 expected in response to climatic change (IPCC, 2001; Dai, 2011, 2013). The 21st century 540 is expected to have an increase in frequency and intensity of droughts, mostly influenced 541 by decrease in rainfall rates and increases in potential evapotranspiration (Dai, 2011, 542 2013; Cook et al., 2014). Evidences of warming at all Brazilian marine ecoregions 543 suggests that most of the Brazilian estuaries are under global warming effects, with this 544 study evidencing that drough events can intensify these changes during the dry seasons 545 (austral winter; Bernardino et al., 2015; Pampuch et al., 2016). The Piraquê-Açú-Mirim 546 estuary is under a marked rainfall oscillation, with more frequent strong rainfall anomaly 547 during in wet seasons. Heavy rainfall, which can cause floods, decrease salinity and 548 temperature, which can affect benthic assemblages (Montagna et al., 2002 Saifullah et al., 549 2016). However, the years of 2015 and 2016 were markedly dryer than if compared to the 550 historical data from the last 60 years in the region, with rainfall volumes 3 to 8 times 551 552 lower than average. This 2015-2016 drought occurred during a very strong El-niño, the highest in this century (> 2.0; ggweather, 2017). El-Niño influences local weather 553 worldwide and mostly leads to increased drought risk (Grimm et al., 1998; Dai, 2013). 554 Warms episods of El-niño are more frequent, persistant and intense since 70', campared 555 to the previous 100 years (IPCC, 2001), which shows the possibility of other strong 556 drought in the coming decades. Drought can be charactezed as a decline in rainfall below 557 the mean monthly precipitation, increases in evapotranspiration, or a combination of the 558 two (Cook et al., 2014). Mean values over a prescribed 30-year interval are sufficient to 559 filter inter-annual fluctuations and anomalies (Hare, 1979). In estuaries, drought seasons 560 may affect estuarine ecosystems by increasing salinity and temperature seasonsdue to the 561 lower fresh water flux from continental drainage and rainfall (McLusky and Elliott, 2004; 562 McLaughlin et al., 2009). The low rainfall during the drought event reduce fresh water 563 presence, increase marine water residence time and evapotranspiration, with resulting 564 salinization within the estuary (Dai, 2013; Saifullah et al., 2016). 565

Water salinity and temperature are related to rainfall changes, with inverse relation 566 with rainfall variations (Palmer et al., 2015). In our study, salinity varied less at the 567 euhaline zones than the polyhaline zones, with higher values in euhaline zones. 568 Temperature was lower in euhaline zones than polyhaline zones at dry and wet seasons 569 of 2015 and Dry season of 2016, and temperature varied more at euhaline zones than 570 571 polyhaline zones. The decrease in salinity range with tidal oscilation in the euhaline zone during the dry season of 2016, and the increase in temperature in euhaline and polyhaline 572 zones across seasons indicate a higher marine water intrusion in Piraquê-Açú-Mirim 573 estuary. During drought seasons water temperature increase, reducing the dissolved 574 oxygen solubility and mix with coastal water (Attrill and Power, 2000; Enquist et al., 575 2003; Stow et al., 2005). Associated with the low rates of oxygen, the high temperature 576 increase oxygen demand by benthic biota, which can promote hypoxia (<2 mg/L), as well 577

hypoxia-derived mortality of the biota (Brown et al., 2004; Harris et al., 2006; VaquerSunyer and Duarte, 2008). The low dissolved oxygen (<1 mg/L) at euhaline zones and</li>
polyhaline zones may influence the decrease in benthic diversity in the Piraquê-AçúMirim estuary at the Dry season of 2016.

Climate change will affect salinity, temperature, turbidity and nutrient availability, 582 583 which can impact primary production (Attrill and Power, 2000; IPCC, 2001). The scarcity of nutrients during low rainfall and drought years can lead to lower Chlorophyll-a 584 variability (Abreu et al., 2010). The drought and El-niño likely influenced the low 585 Chlorophyll-a and pheopigments at both euhaline and polyhaline zones across all seasons 586 in the Piraquê-Açú-Mirim estuary (Attrill and Power, 2000; Drake et al., 2002; Grilo et 587 al., 2009; Abreu et al., 2010). Drought may act differently in each estuary, ranging from 588 major influences, as ENSO, to regional changes in precipitation and evapotranspiration. 589 Particle sedimentation in estuaries depend on aggregation dynamics, mostly influenced 590 by particle concentration, size, fluid dynamics and salinity (Eisma 1986; Mari 2008; Mari 591 et al., 2012). Droughts can change sedimentation in estuaries generally increasing silt 592 fractions of sediment towards the mouth of an estuary (MacKay et al., 2010; Mari et al., 593 2012; Dittmann et al., 2015). Additionally, salinization and pH increases during droughts 594 modify the aggregation processes of adsorption and flocculation, which may promote fast 595 sinking of aggregates (Eisma, 1986; Komar, 1996; Mari et al., 2012). The marine water 596 intrusion in polyhaline zones in the Piraquê-Açú-Mirim estuary likely contributed to the 597 increase of mud content at both euhaline zones and polyhaline zones in the Piraquê-Açú-598 Mirim estuary at the Dry season of 2016. 599

Benthic estuarine assembages exhibit changed patterns of distribution, abundance 600 and structure during climatic events (Drake et al., 2002; Pillay and Perissinotto, 2008; 601 602 Pollack et al., 2011; Medeiros et al., 2016). These changes reorganize ecosystem functions on regional and/or local scales, which can couse trophic cascade (Folke et al., 2004). 603 Salinity is an important factor controling benthic estuarine assemblages, estuarine 604 environments acts as a filter, supporting species with functional or phenotypic 605 characteristics suitable for establishment and persistence in each haline zone (Pillay and 606 Perissinotto, 2008; Barros et al., 2012; Mariano and Barros, 2015). The power and 607 frequency increase of drought events can influence benthic assemblages by changes in 608 sediment and physiological stress, with can results in changes in biogeochemical process 609 and ecosystem functions (Elliott and Whitfield, 2011; Mcleod et al., 2011; Dittmann et 610 al., 2015; Montagna and Palmer, 2012). Grain size is a fundamental influence to 611 macrofaunal composition in soft sediments (Carvalho et al., 2017). The increase in 612 sediment complexity usually contain higher diversity of macrofaunal taxa, while the 613 reduction in grain size decrease the space between sediment, where benthic biota lives 614 (Gray and Elliott, 2009). The decrease in grain size reduce permeability, oxygen 615 penetration depth and nutrient rates (Pratt et al., 2014). Increase in mud content decrease 616 macrofaunal density and species richness (Thrush et al., 2003; Anderson 2008), which 617 affects ecological functions, due changes in benthic structure and loss of key species 618 (Pratt et al., 2014). In the present study mud content was high during all seasons, with 619 increase in dry season of 2016, that influences seasonal changes on macrofaunal 620 621 assemblages (total biomass, Richness, Pielou J 'and Shannon H') of Piraquê-Açú-Mirim estuary. Temperature of water has been reported as being a major controlling variable for 622 species, which could affect the metabolism, growth and reproduction (Attrill and Power, 623 2010; Bishop et al., 2006; Dolbeth et al., 2011; Munguia et al., 2013). The increase in 624 temperature of euhaline and polyhaline zones at Dry 2016 compared to Dry 2015 and Wet 625 2015 influences the decrease in Macrofaunal Richness and increase in Pielou J', while 626 salinity did not influence macrofaunal changes of Piraquê-Açú-Mirim estuary. 627

Similar habitats (e.g. healthy mudflats within the same haline zone) should have 628 a pool of similar species (both functionally and morphologically) and may reflect local 629 environmental filters that regulate the dispersion of species along haline gradients 630 (Medeiros et al., 2016). Salinity is considered the most important variable to macrofaunal 631 assemblages estructure in estuaries, but other variables (e.g. temperature and grain size) 632 633 can be more important to these changes, which evidence that macrofaunal assemblages are not effected in a simple or linear way (Teske and Wooldridge, 2003; MacKay et al., 634 2010; Palmer et al., 2015; Medeiros et al., 2016). Drought events can lead to marked 635 changes in estuarine assemblages (IPCC, 2001; Pillay and Perissinotto, 2008, 2009; 636 MacKay et al., 2010; Dittmann et al., 2015; González-Ortegón et al., 2015; Table 5, 637 references therein). High presence of marine opportunistic taxa, including Capitellidae 638 and Oligochaete evidence salinity intrusion and changes in macrofaunal composition 639 (Hastie and Smith, 2006; Pillay and Perissinotto, 2008; MacKay et al., 2010; Table 5, 640 references therein). Euhaline and polyhaline zones of Piraquê-Açú-Mirim estuary were 641 dominated by the polychaets Magelonidae, Sternaspidae and Capitellidae, with shifts in 642 taxa composition (Oligochaete and Paraonidae). Salinity of water was not related to any 643 macrofaunal assemblages parameter across seasons, wheras temperature was related to 644 macrofaunal richness and Pielou J', and taxa distribution (e.g. Phyllodocidae and 645 Cirratulidae) in euhaline and polyhaline zones of Piraquê-Açú-Mirim estuary. 646

Brazil has 8,000 km of coast line with over one hundred estuaries from the tropical 647 equator in the North to South lagoons, with a general lack of long term ecological 648 649 monitoring programs (Bernardino et al., 2016). Geomorphological and oceanographic conditions changes on Brazilian estuaries are evident, where dominated by riverine inputs 650 are more frequent in the N and NE, while bays, drowned estuaries and lagoons in SE and 651 S (Dominguez, 2006; Bernardino et al., 2016). This monitoring is one of the most 652 extensive studies ever carried within Brazilian estuaries, totalizing 14 months of 653 monitoring (May 2015 to June 2016; Bernardino et al., 2015, 2016). The temporal 654 analysis of euhaline and polyhaline zones of Piraquê-Açú-Mirim estuary provided 655 significant advance in the understanding of natural climatic oscillations on benthic 656 assemblages of east Brazilian estuaries. The system recovery can be faster to especifc 657 permanents changes, depending to the intensity of the event and helth of the estuary. All 658 of these evidence the necessity to understand the impacts coused by droughts and floods 659 from others impacts (natural or antropogenic; IPCC, 2001; Worm et al., 2006; Table 5, 660 references therein). However, long-term assessment of naturals oscillation, as droughts, 661 662 elucidated patterns changes and be usefull to overall management, providing more accurate models (Dolbeth et al., 2011; Elliott and Whitfield, 2011; Mcleod et al., 2011; 663 Bernardino et al., 2016). 664

665 Our results corroborate that temperature influence benthic shifts during drought. 666 Despite faunal composition shifts, macrofaunal density, biomass, richness, H' and J' 667 decrease, which evidence a decrease in diversity and dominant taxa density. Studies still 668 needed to better understand the relationship between weather-dependent variables and 669 benthic assemblages, as well factors controlling their variability.

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### 671 **Conflict of interest**

The authors declare no actual or potential conflict of interest.

#### 674 **Contribution**

- LEOG participated in sampling and analyzed data. LEOG and AFB wrote the manuscript.
- All authors have approved the final article.
- 677

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- 1023

1025 **Table captions**Table 1. Changes of pH, Total of dissolved solids (TDS) and Dissolved oxygen (mg/L, DO) in euhaline and polyhaline zones between dry and wet seasons of 2015 and dry season of 2016 at Piraquê-Açú-Mirim estuary.

		Euhalin	e zone		Polyhal	ine zone	
		pН	TDS	DO	Ph	TDS	DO
	May 2015	-	-	-	-	-	-
Dryl	June 2015	-	-	-	-	-	-
	July 2015	-	-	-	-	-	-
	August 2015	8.7±0	25.1±1.4	5.4±0	8.7±0	26.7±0.1	5.8±0.3
	September 2015	7.8±0	27.8±0	3.9±0.3	8.1±0	25.4±0	3.2±0.2
	October 2015	8.2±0	29.1±0	3±0.2	8.3±0.1	28±0.1	2±0.1
	November 2015	8±0.1	29.6±0	1.9±0	8.3±0.1	29.2±0	1.2±0.2
Wet	December 2015	8±0	29.1±0.1	6.4±0.2	7.6±0	27±0	3.9±0.2
2015	January 2016	8.1±0.1	24.9±0.2	3.2±0.7	7.7±0	17.8±0.2	3.3±0
	February 2016	8.8±0.4	28±0.1	1.7±0.1	7.8±0	25.8±0.2	1.4±0
	March 2016	8.4±0	29.1±0.1	2.4±0.5	8.0±0.1	27.6±0.1	1±0.1
	April 2016	8.3±0	29.4±0	0.7±0.3	7.9±0	27.8±0.2	0.6±0
Dry2	May 2016	8.9±0.2	30.2±0	<lq< td=""><td>8±0</td><td>29.8±0</td><td><lq< td=""></lq<></td></lq<>	8±0	29.8±0	<lq< td=""></lq<>
	June 2016	9.1±0.3	30.5±0	<lq< td=""><td>8.3±0</td><td>30±0.1</td><td><lq< td=""></lq<></td></lq<>	8.3±0	30±0.1	<lq< td=""></lq<>

	Euhaline zone						Polyhaline zon	е				
Taxa	Dry 2015		Wet season		Dry 2016		Dry 2015		Wet season		Dry 2016	
	Density (SD)	Rel %	Density (SD)	Rel %	Density (SD)	Rel %	Density (SD)	Rel %	Density (SD)	Rel %	Density (SD)	Rel %
Magelonidae P \$	1202 (918)	47.8	619.8 (593.9)	52.1	253.5 (426.2)	35.6	885.4 (675.8)	47.9	576.4 (655.9)	58.0	270.8 (282.7)	43.3
Spionidae P	64.6 (148.2)	2.6	1.7 (14.7)	0.1	6.9 (29)	1.0	52.1 (87)	2.8	10.4 (62.1)	1.0	13.9 (49.8)	2.2
Sternaspidae P \$	800 (868.2)	31.8	288.2 (432.8)	24.2	66 (121.3)	9.3	360.4 (387.9)	19.5	158 (229.8)	15.9	41.7 (79.1)	6.7
Cossuridae P	4.2 (22.6)	0.2	6.9 (28.8)	0.6	20.8 (70)	2.9	12.5 (37.8)	0.7	17.4 (60.5)	1.7	13.9 (49.8)	2.2
Nereididae P	29.2 (962.5)	1.2	6.9 (28.8)	0.6	34.7 (97)	4.9	58.3 (106.6)	3.2	8.7 (32)	0.9	3.5 (20.8)	0.6
Capitellidae P \$	112.5 (192.1)	4.5	95.5 (228.4)	8.0	163.2 (233)	22.9	137.5 (211.8)	7.4	85.1 (177.8)	8.6	97.2 (140.5)	15.6
Goniadidae P \$	108.3 (140.9)	4.3	71.2 (86.1)	6.0	34.7 (64.1)	4.9	95.8 (128.9)	5.2	50.4 (95.4)	5.1	66 (101.3)	10.6
Pilargidae P	16.7 (42.9)	0.7	15.6 (51.1)	1.3	31.3 (75.4)	4.4	18.8 (64.4)	1.0	6.9 (28.8)	0.7	13.9 (49.8)	2.2
Paraonidae P \$	62.5 (101.6)	2.5	12.2 (52.1)	1.0	0.0 (0.0)	0.0	145.8 (167.8)	7.9	29.5 (57.4)	3.0	0.0 (0.0)	0.0
Cirratulidae P	2.1 (16.1)	0.1	1.7 (14.8)	0.1	0.0 (0.0)	0.0	10.4 (41.8)	0.6	5.2 (25.2)	0.5	3.5 (20.8)	0.6
Lumbrineridae P	18.8 (68.4)	0.7	0.0 (0.0)	0.0	0.0 (0.0)	0.0	2.1 (16.1)	0.1	0.0 (0.0)	0.0	0.0 (0.0)	0.0
Amphinomidae P	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0	2.1 (16.1)	0.1	0.0 (0.0)	0.0	0.0 (0.0)	0.0
Scalibregmatidae P	18.8 (45)	0.7	15.6 (55.3)	1.3	3.5 (20.8)	0.5	2.1 (16.1)	0.1	0.0 (0.0)	0.0	0.0 (0.0)	0.0
Phyllodocidae P	18.8 (60.1)	0.7	3.5 (20.7)	0.3	0.0 (0.0)	0.0	8.3 (31.4)	0.5	0.0 (0.0)	0.0	0.0 (0.0)	0.0
Sabellidae P	0.0 (0.0)	0.0	0.0 (0.0)	0.0	3.5 (20.8)	0.5	0.0(0.0)	0.0	0.0 (0.0)	0.0	13.9 (83.3)	2.2
Sabellariidae P	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0(0.0)	0.0	0.0 (0.0)	0.0	3.5 (20.8)	0.6
Oligochaete O	0.0 (0.0)	0.0	0.0 (0.0)	0.0	41.7 (126.8)	5.9	0.0(0.0)	0.0	0.0 (0.0)	0.0	48.6 (153.2)	7.8
Gammaridae C	2.1 (16.1)	0.1	0.0 (0.0)	0.0	0.0 (0.0)	0.0	8.3 (31.4)	0.5	0.0 (0.0)	0.0	0.0 (0.0)	0.0
Ocypoddidae C	4.2 (22.6)	0.2	5.2 (32.8)	0.4	17.4 (53.1)	2.4	2.1 (16.1)	0.1	3.5 (20.7)	0.3	0.0 (0.0)	0.0
Atyidae C	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0(0.0)	0.0	0.0 (0.0)	0.0	3.5 (20.8)	0.6
Penaeidae C	10.4 (34.9)	0.4	0.0 (0.0)	0.0	0.0 (0.0)	0.0	2.1 (16.1)	0.1	0.0 (0.0)	0.0	0.0 (0.0)	0.0
Cironalidae C	0.0 (0.0)	0.0	0.0 (0.0)	0.0	10.4 (62.5)	1.5	0.0(0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0
Alpheidae C	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0	3.5 (20.8)	0.6
Cumacea C	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0(0.0)	0.0	0.0 (0.0)	0.0	3.5 (20.8)	0.6
Vitrinellidae M	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0	10.4 (74.9)	1.0	0.0 (0.0)	0.0
Corbiculidae M	4.2 (22.6)	0.2	17.4 (76.5)	1.5	10.4 (76.5)	1.5	0.0 (0.0)	0.0	10.4 (45.7)	1.0	3.5 (20.8)	0.6
Solecurtidae M	0.0 (0.0)	0.0	10.4 (34.8)	0.9	3.5 (20.8)	0.5	0.0 (0.0)	0.0	13.9 (49.4)	1.4	6.9 (29)	1.1
Mytilidae M	0.0 (0.0)	0.0	3.5 (20.8)	0.3	3.5 (20.8)	0.5	0.0(0.0)	0.0	3.5 (20.7)	0.3	0.0 (0.0)	0.0

1029Table 2. Density (ind.m²) and Relative abundance (%) of macrofaunal assemblages in euhaline and polyhaline zones between dry and wet seasons of 2015 and dry season of10302016 at the Piraquê-Açú-Mirim estuary. \$ = top ranked macrofaunal species. P = Polychaete, O = Oligochaete, M = Mollusk, C = Crustacean.

Nuculidae M	6.3 (27.5)	0.2	0.0 (0.0)	0.0	0.0 (0.0)	0.0	2.1 (16.1)	0.1	0.0 (0.0)	0.0	0.0 (0.0)	0.0
Tellinidae M	14.6 (46.6)	0.6	6.9 (35.7)	0.6	0.0 (0.0)	0.0	14.6 (40.5)	0.8	3.5 (20.8)	0.4	0.0 (0.0)	0.0
Nemertea	12.5 (59.6)	0.5	6.9 (35.7)	0.6	6.9 (29)	1.0	25 (55.4)	1.4	0.0 (0.0)	0.0	13.9 (65.3)	2.2
Sipuncula	0.0 (0.0)	0.0	0.0 (0.0)	0.0	0.0 (0.0)	0.0	4.2 (22.6)	0.2	0.0 (0.0)	0.0	0.0 (0.0)	0.0

1032 1033 Table 3. Distance-based linear model (DistLM) of Bray-Curtis similarities on macrofaunal assemblages and sedimentary properties (Mud content, Total organic content, Calcium

3	carbonate, Chlorop	hyll- <i>a</i> and P	heopigm	nents)	between o	lry and	d wet seasons o	of 2015 an	d dry seas	on of 2016 ir	n the Piraqu	ıê-Açú-Mirin	n estuary	. Bold =	p significant.	
	•	a	<b>a</b> ()	<b>D</b>	1 5			D								

n=3	SS(trace)	Pseudo-F	р	Prop.	
Mud content	13629	13.392	0.001	0.10853	
Total organic content	1827.6	1.6247	0.135	1.46E-02	
Calcium carbonate	1390	1.2313	0.278	1.11E-02	
Chlorophyll-a	2007	1.7867	0.095	1.60E-02	
Pheopigments	1442.2	1.2781	0.259	1.15E-02	
res.df: 110					

Table 4. Best correlations of macrofaunal assemblages and sedimentary properties (Mud content, Total organic content, Calcium carbonate, Chlorophyll-*a* and Pheopigments) in the Piraquê-Açú-Mirim estuary.  $P_W$ -Weighted Spearman coefficients. Mud = Mud content, TOC = Total organic content, CaCO3 = Calcium 

carbonate, $Chl-a = Chl$	orophyll- <i>a</i> a	nd Pheo = Pheopigments.
No. of variables	Pw	Variables
1	0.107	Mud
2	0.103	Mud, CaCO3
3	0.066	Mud, TOC, CaCO3
3	0.061	Mud, CaCO3, Chl-a
1	0.049	CaCO3
2	0.047	Mud, TOC
2	0.04	Mud, Chl-a
3	0.038	Mud, CaCO3, Pheo
2	0.033	Mud, Pheo
4	0.032	Mud, TOC, CaCO3, Chl-a

Tuble 5. Bites, country und	durution of the	monitoring program with drought events in estuaries compared to	I maque I n	a minin ostaary. rota	a rotar sampning events:
Sites	Contry	Duration of the monitiring	Total	Benthic invertebrate	References
Guadalupe estuary	United States	2 years, 3 months first and 2 second year	5	Macrofauna	Montagna and Kalke 1992
Guadalupe estuary	United States	2 years, 3 months first and 2 second year	5	Meiofauna	Montagna and Kalke 1992
Nueces estuary	United States	2 years, 3 months first and 2 second year	5	Macrofauna	Montagna and Kalke 1992
Nueces estuary	United States	2 years, 3 months first and 2 second year	5	Meiofauna	Montagna and Kalke 1992
Nueces estuary	United States	5 years, 4 months/year	20	Macrofauna	Montagna et. al. 2002
Nueces estuary	United States	5 years, 4 months/year	20	Meiofauna	Montagna et. al. 2002
Nueces estuary	United States	8 weeks	-	Macrofauna	Ritter et al., 2005
Nueces estuary	United States	30 years, 1968-1970 and 1997-1998	10	Macoma spp.	Bishop et al., 2006
Nueces estuary	United States	Over 14 years	25-114	Macrofauna	Montagna et. al. 2009
Lavaca-Colorado Estuary	United States	~20 years, Quarterly from April 1988 to October 2008	139	Macrofauna	Pollack et al., 2011
Lavaca-Colorado estuary	United States	>15 years, montly	~180	Macrofauna	Montagna and Palmer 2012
Lavaca-Colorado estuary	United States	>15 years, montly	$\sim \! 180$	Epifauna	Montagna and Palmer 2012
Guadalupe estuary	United States	>15 years, montly	$\sim \! 180$	Macrofauna	Montagna and Palmer 2012
Guadalupe estuary	United States	>15 years, montly	$\sim \! 180$	Epifauna	Montagna and Palmer 2012
Nueces estuary	United States	>15 years, montly	~180	Macrofauna	Montagna and Palmer 2012
Nueces estuary	United States	>15 years, montly	~180	Epifauna	Montagna and Palmer 2012
Nueces Estuary	United States	2 years, different sampling	-	Macrofauna	Montagna et. al. 2015
Nueces Estuary	United States	2 years, different sampling	-	Epifauna	Montagna et. al. 2015
Lavaca-Colorado estuary	United States	>15 years, between 1987 and 2009	-	Macrofauna	Palmer and Montagna 2015
Guadalupe estuary	United States	>15 years, between 1987 and 2009	-	Macrofauna	Palmer and Montagna 2015
Nueces estuary	United States	>15 years, between 1987 and 2009	-	Macrofauna	Palmer and Montagna 2015
Caloosahatchee Estuary	United States	10 years, February 1986 to April 1989, and October 1994 to December 1995	-	Macrofauna	Palmer et al., 2015
Gulf of Lion	France	>30 years, 1963, 1969, 1972, 1975, 1976 and then every year from 1984 to 1986 and 1993 to 1996	12	Polychaets	Salen-Picard and Arlhac 2002
Rhone river delta	France	3 years, 3 months/year	12	Macrofauna	Salen-Picard et al., 2003
Mondego estuary	Portugal	>10 years, fortnightly in first 18 months, monthly thereafter	~130	Macrofauna	Dolbeth et al., 2007
Mondego estuary	Portugal	13 years, January 1993 to January 1997, and January 1999 to December 2005, monthly	~130	Hydrobia ulvae	Cardoso et al., 2008
Mondego estuary	Portugal	>15 years, January 1993 to September 1995; March 1999 to November 2008, first 18 monthsfortnightly and monthly	~150	AmphipodS	Grilo et al., 2009

1040 Table 5. Sites, country and duration of the monitoring program with drought events in estuaries compared to Piraquê-Açú-Mirim estuary. Total = Total sampling events.

		thereafter			
Mondego estuary	Portugal	8 years, spring tide of 1990, 1992, 1998, 2000, 2002, 2004 to 2006	-	Macrofauna	Neto et al., 2010
Mondego estuary	Portugal	14 years, fortnightly/first 18 months and monthly thereafter	186	Macrofauna	Dolbeth et al., 2011
	-	>15 years, January 1993 to September 1995; March 1999 to			
Mondego estuary	Portugal	November 2008, first 18 months fortnightly and monthly thereafter	~150	Macrofauna	Grilo et al., 2011
Mondego estuary	Portugal	5 year, 2004 to 2008	-	Macrofauna	Veríssimo et al., 2012
Mondego estuary	Portugal	5 year, 2004 to 2008	-	Macrofauna	Veríssimo et al., 2013
Mondego estuary	Portugal	10 years, February 1993 to June 1994 (fortnightly) and monthly until 2002	137	Macrofauna	Dolbeth et al., 2014
Mondego estuary	Portugal	7 years, monthly from January 1999 to December 2005	84	Scrobicularia plana	Verdelhos et al., 2014
Guadalquivir estuary	Spain	2 years, monthly from May 1997 to April 1999	24	Hyperbenthos	Drake et al., 2002
Guadalquivir estuary	Spain	9 years, monthly from May 1997 to January 2005	24	Shrimps	González-Ortegón <i>et al.,</i> 2006
Guadalquivir river basin	Spain	12 years, monthly	140	Macrofauna	González-Ortegón <i>et al.,</i> 2015
Hawkesbury estuary	Australia	2 years, 3-4 months/year	7	Macrofauna	Jones 1990
Brunswick estuary	Australia	2 years, December 2000 to December 2002, sampled 1-2 months	16	Macrofauna	Eyre and Ferguson 2005
Brunswick estuary	Australia	2 years, December 2000 to December 2002, sampled 1-2 months	16	Meiofauna	Eyre and Ferguson 2005
Hopkins river estuary	Australia	2 years, 3 times in 1997 and 1 in 1998	4	Soletellina alba	Matthews 2006
Coorong lagoons	Australia	10 years, 1 sample/years	10	Macrofauna	Dittmann et al., 2015
Murray Mouth	Australia	10 years, 1 sample/years	10	Macrofauna	Dittmann et al., 2015
Nhlabane estuary	South Africa	3 years, 1 sample at intervals of three months	12	Macrofauna	Mackay and Cyrus 2001
Siyaya estuary	South Africa	3 years, 1 sample at intervals of three months	12	Macrofauna	Mackay and Cyrus 2001
Santa Lucia estuary	South Africa	1 year, 4 months	4	Macrofauna	Pillay and Perissinotto 2008
Santa Lucia estuary	South Africa	2 year, 4 months	5	Meiofauna	Pillay and Perissinotto 2009
Santa Lucia estuary	South Africa	5 years, 2 months/year, 1 sample in 2004 and 2005	8	Macrofauna	MacKay et al., 2010
Santa Lucia estuary	South Africa	Laboratory experiment with estuarine water	-	Solen cylindraceus	Nel et al., 2011
Santa Lucia estuary	South Africa	3 years, quarterly from August 2006 to May 2008	7	Meiofauna	Bownes and Perissinotto 2012
Santa Lucia estuary	South Africa	4 years	-	Macrofauna	Pillay and Perissinotto 2013
Piraquê-Açú-Mirim estuary	Brazil	1.5 years, 14 months, monthly	14	Macrofauna	This study

## 1042 Figure captions





Figure 1. Map of the Piraquê-Açú-Mirim estuary indicating the locations of sampling sites.



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Figure 2. Upper: Mean annual rainfall from 1948 to 2016 (ANA and UFES weather stations; 1948 to 2016) 1047 at Piraquê-Açú-Mirim estuary; and Bottom: monthly total rainfall during this study (ANA and UFES 1048 weather stations; 2015-2016) and background information (ANA weather stations; 1948 to 2014) from 1049 Piraquê-Açú-Mirim. Estuary.



1051 Figure 3. Spring and neap tides salinity in the monitoring days of Dry and Wet seasons of 2015 and Dry season of 2016 seasons in euhaline and polyhaline zones at Piraquê-

Açú-Mirim estuary. Monitoring days at Euhaline zones (40 days Dry season of 2015 (6d), Wet season of 2015 (24d) and Dry season of 2016 (10d)); and Polyhaline zones (35 days; Dry season of 2015 (5d), Wet season of 2015 (22d) and Dry season of 2016 (8d)).



Figure 4. Spring and neap tides temperature (°C) in the monitoring days of Dry and Wet seasons of 2015 and Dry season of 2016 seasons in euhaline and polyhaline zones at

1056 Piraquê-Açú-Mirim estuary. Monitoring days at Euhaline zones (40 days Dry season of 2015 (6d), Wet season of 2015 (24d) and Dry season of 2016 (10d)); and Polyhaline zones (35 days; Dry season of 2015 (5d), Wet season of 2015 (22d) and Dry season of 2016 (8d)).



1058SeasonsSeasons1059Figure 5. Macrofaunal Density (ind.m²), Biomass (g.m²), Richness, Diversity H' and Pielou J' in euhaline zones and polyhaline zones of Dry and Wet seasons of 2015 and Dry1060season od 2016 at Piraquê-Açú-Mirim estuary.



Figure 6. Non-metric multidimensional scaling (MDS) plot of macrofaunal assemblages in euhaline and polyhaline zones at Piraquê-Açú-Mirim estuary.



Figure 7. Distance-based linear model plot (dbRDA) using the sedimentary properties to explain macrofaunal assemblages changes in Piraquê-Açú-Mirim estuary. Mud = Mud content, TOC = Total organic content, CaCO3 = Calcium carbonate, Chl-a = Chlorophyll-a and Pheo = Pheopigments.

## III. CONCLUSÃO

Nossos resultados corroboram que a temperatura influencia mudanças em assembleias bentônicas durante eventos de seca. No estuário do Piraquê-Açú-Mirim temperatura da água influenciou a estrutura da macrofaunal bentônica (riqueza S e Pielou J'), e alterações da composição dos táxons das assembleias bentônicas (Phyllodocidae and Cirratulidae);

Globalmente são escassos os monitoramentos estuarinos, especialmente nos países subdesenvolvidos e em desenvolvimento. Este monitoramento é um dos estudos mais extensos já realizados nos estuários brasileiros, totalizando 14 meses de monitoramento, proporcionando um avanço significativo na compreensão das oscilações climáticas naturais em assembleias bentônicas em estuários do leste brasileiro;

Ressaltamos a necessidade de monitoramentos de longa duração, como os aqui exemplificados (> 10 anos), para entender as respostas estuarinas aos efeitos de seca e inundação. Este conhecimento possibilitará melhor gerir estuários importantes, como Piraquê-Açú-Mirim e as baias de Todos os Santos e Paranaguá, frente as previsões de aumento da intensidade e frequência de secas durante as mudanças climáticas no século atual.

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## V. Anexos

Em anexos são apresentadas as normas de submissão da Revista "Marine Environmental Research".



## MARINE ENVIRONMENTAL RESEARCH

### **AUTHOR INFORMATION PACK**

#### **TABLE OF CONTENTS** Marine Environmental Description • p.1 Research Audience p.1 • **Impact Factor** p.1 • Abstracting and Indexing p.2 • **Editorial Board** p.2 • **Guide for Authors** p.3 ISSN: 0141-1136

### DESCRIPTION

*Marine Environmental Research* publishes original research papers on chemical, physical, and biological interactions in the **oceans** and **coastal waters**. The journal serves as a forum for new information on biology, chemistry, and toxicology and syntheses that advance understanding of **marine environmental** processes.

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– The biochemical, physiological, and ecological consequences of contaminants to marine organisms and ecosystems

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- Methodological papers describing improved quantitative techniques for the marine sciences.

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2015: 2.769 © Thomson Reuters Journal Citation Reports 2016

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